



Efficient adsorption capability of banana and cassava biochar for malachite green: Removal process and mechanism exploration

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Abstract

In this study, banana biochar (BB) and cassava biochar (CB) were roasted in a tube furnace at 400°C for 2 h, and their adsorption effects on malachite green (MG) were studied. The removal of MG with respect to the solution concentration and reaction time was also investigated. The results show that the optimal amount of biochar is 2 g/L. Studies of the adsorption kinetics and adsorption isotherms show that the pseudo-second-order kinetics can most accurately reflect the adsorption process. Freundlich model fits the experimental data well. The theoretical saturated adsorption capacities of BB and CB are 1,092.80 mg/g and 261.42 mg/g, respectively. SEM-EDS, TGA, BET, FTIR, XRD, element analysis and other characterization techniques were used to explore the adsorption mechanism. Based on the characterization results, it is speculated that the adsorption of BB and CB for MG mainly includes chemical adsorption and physical adsorption.

Keywords: Adsorption, Adsorption mechanisms, Biochar, Biomass materials, *Malachite Green*



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1. Introduction

With economic development and industrialization, the types and output of synthetic dyes have rapidly increased. According to statistics, there are more than 10,000 types of dyes worldwide, and the annual output exceeds 700,000 tons [1]. The data show that during processing and discharge, dyes cause serious pollution. In the processing stage, the dye output may decrease by 30%, and the amount of dye that is directly discharged into wastewater without treatment can reach 10-15% [2]. Most dyes have high colour, toxicity and carcinogenicity, which make them harmful to human health. In addition, dye wastewater will destroy the aquatic ecosystem, affect the *chemical oxygen demand* (COD) and *biological oxygen demand* (BOD) in water, and reduce the light transmission capacity of water [3].

Malachite green (MG) is a common cationic dye and easily soluble in water. MG can kill fungi, bacteria and parasites in water and dye materials such as textiles, wool, and paper products [4]. Therefore, it is widely used in the fishery and textile industry. MG is also a highly toxic triphenylmethane (TMP) teratogen and carcinogen and has been classified into the second category of dangerous goods. However, many countries continue using MG in large quantities because of its low cost and fast results [5]. Therefore, many scholars have begun to focus on the removal of MG in wastewater.

At present, the number of treatment technologies for MG in wastewater is continuously increasing, including coagulation, precipitation, catalysis, photocatalysis, and biodegradation [6]. However, coagulation and precipitation are often too costly, the biodegradation operation is complex, and microorganisms can easily die due to changes in culture conditions [7]. Most dyes are resistant to biodegradation because of their structural complexity [8]. Moreover, biological

1 processes will produce a large amount of sludge, which easily causes new pollution. Catalysis
2 and photocatalysis may also generate new toxic substances, even if they can degrade dyes and
3 fundamentally remove organic pollutants [9]. In contrast, the adsorption method has simple
4 operation, a low cost, and high efficiency and has become one of the most commonly used
5 methods to treat dye wastewater [10]. In recent years, research on the use of agricultural waste to
6 prepare biochar as an adsorbent has gradually increased; example wastes include bamboo [11],
7 crab shell [12], cactus [13], and bagasse [14].

8 Biomass is the general name for organic matter that is directly or indirectly derived from
9 various green plants, such as agricultural waste (chaff, fruit husks, crop straw, etc.), forestry
10 waste (grass, branches, sawdust, etc.), aquatic plants (water hyacinth, algae, etc.), municipal
11 garbage and sewage sludge [15]. Biochar is a type of porous carbon-rich solid material prepared
12 by the pyrolysis of biomass under oxygen-free and anoxic conditions. As a new multifunctional
13 material, biochar is widely used in water pollution control, soil fertility improvement, soil carbon
14 fixation, soil remediation, microbial fuel cell electrodes and other fields [16]. In recent years,
15 biochar prepared from crop straw has attracted much attention due to its outstanding effect. It has
16 large specific surface area, complex pore structure and abundant surface active functional groups,
17 and it has great potential as an adsorbent. However, the physical and chemical properties of
18 biochar largely depend on the type of raw materials and production conditions. Therefore, from
19 the perspective of biochar production, there are two important aspects: selecting the appropriate
20 raw materials and understanding the key production parameters that affect the physical and
21 chemical properties of biochar [17]. In addition, to realize the benefits of biochar in agriculture,

1 it is necessary to design a simple, easy-to-use and economically feasible technology to produce
2 biochar from ordinary biomass waste, which farmers can use to prepare biochar themselves.

3 Guangxi is the province with the most cassava and banana cultivation in China. Every
4 year, a large amount of cassava and banana straw waste is generated, which causes waste
5 disposal problems. In this paper, cassava and banana straw are used as raw materials, and low-
6 cost and high-adsorption biochars are made by a simple pyrolysis process without further
7 modification and used to adsorb the typical cationic dye MG. This approach can solve the
8 problem of waste disposal. (1) The batch adsorption of MG by banana biochar (BB) and cassava
9 biochar (CB) was studied to analyse the adsorption kinetics, adsorption isotherm, pH value, and
10 effect of dosage on dye adsorption. (2) Additionally, the physical and chemical properties of BB
11 and CB before and after adsorption were studied by characterization to reveal their adsorption
12 mechanism.

13

14 **2. Materials and Methods**

15 **2.1. Materials**

16 The banana and cassava straws were obtained from the Guilin Qixing Farmers Market, and MG
17 (molecular formula: $C_{52}H_{54}N_4O_{12}$; structural formula in Fig. 1; $\lambda_{max} = 618$ nm) was purchased
18 from Guilin Bell Experimental Equipment Co., Ltd. The reagents were of analytical grade.

19

20 **2.2. Preparation of BB and CB**

21 The cassava straw was cut into small pieces, simply dried in the sun, placed in an oven to dry at
22 60-65°C, ground with a grinder, and passed through a standard sieve of 0.425 mm to obtain

1 cassava straw powder. The accurately weighed 3 g of tapioca straw powder was placed in a
2 vacuum/atmosphere tube furnace, the N₂ flow rate was set to 500 mL/min, the heating rate was
3 set to 10°C/min, and the sample was kept at 400°C for carbonization under anoxic conditions for
4 2 h. After the furnace was cooled to room temperature, the sample was removed and ground in a
5 crucible. The resulting CB was passed through a 0.425-mm standard sieve and sealed for
6 preservation. Since banana straw contains a large amount of water, water was first removed by a
7 juicer; then, the straw was dried. Subsequent steps were consistent with those for CB preparation,
8 and the resultant product was named BB.

10 **2.3. Characterization of BB and CB**

11 The surface morphological characteristics of the biochars were observed by scanning electron
12 microscopy (SEM, FEI Inspect F 50, USA). The surface functional groups of the adsorbents
13 were analysed by Fourier transform infrared spectroscopy (FTIR, Nicolet 380, USA). X-ray
14 diffractometry (XRD, Panalytical, Netherlands) was used to analyse the mineral composition and
15 crystal structure of the biochars. The porosity and specific surface area were determined by
16 Brunauer-Emmett-Teller (BET) method (ASAP24920, USA). Thermogravimetric analysis (TGA,
17 LabSys Evo TG-DTA, TA Q600, USA) was used to determine the stability of the biochars. The
18 element content of the adsorbents was determined by an elemental analyser (Vario EL III,
19 Germany). The TOC content before and after adsorption was determined by TOC analyzer (TOC,
20 V_{CPH}, Japan)

22 **2.4. Batch Adsorption Experiments**

1 The adsorption properties of BB and CB for MG were compared under different conditions, and
2 the adsorption isotherm and adsorption kinetic models were fitted. The details are as follows:
3 different doses of adsorbent were accurately weighed in a 100-mL volumetric flask, and 50 mL
4 MG solution of different initial concentrations was added. The mixture was oscillated in an
5 oscillator at 150 rpm for different times. After the adsorption was completed, the mixture was
6 filtered with a 0.45- μ m microporous membrane and analysed with a 722 S spectrophotometer at
7 a wavelength of 618 nm. The adsorption parameters were as follows: the adsorbent dose was 0.5-
8 3.5 g/L, the initial concentration of MG was 500-1,500 mg/L for BB and 300-800 mg/L for CB,
9 the oscillation time was 1-1,448 min, and the pH was 2-8. All experiments were performed at
10 room temperature.

11 Calculation of the adsorption amount:

$$12 \quad Q_e = (C_0 - C_e) * V / m \quad (1)$$

13 where Q_e (mg/g) is the biochar adsorption amount; C_0 (mg/L) and C_e (mg/L) are the
14 concentrations of MG before and after adsorption, respectively; V (L) is the volume of the
15 adsorption solution; m (g) is the amount of adsorbent.

16

17 **3. Results and Discussion**

18 **3.1. Material Characterization Analysis**

19 3.1.1. SEM analysis

20 The SEM analysis results of the two biochars before and after adsorption are shown in Fig. 1.
21 Before adsorption, BB has a rough, fibrous cylindrical surface, while CB is in an irregular state.
22 Under the carbonization condition of 400°C, the surface of the two adsorbents becomes rough

1 with irregular random porous shapes due to the elimination of volatile matter. The pore size of
2 BB changes after adsorption, and small balls are generated on the surface, while the inner pore
3 size of CB increases after adsorption, which is consistent with the pore size analysis results. It is
4 speculated that the filling of MG may make the pore diameter smaller and prone to expansion,
5 and the production of pellets may be accompanied by the formation of new substances.

6

7 3.1.2. Element analysis

8 Table 1 shows the results of biochar element analysis. The main components of biochar are C, H,
9 N and O. The H/C ratio is used to indicate the aromaticity of biochar, (O+N)/C indicates the
10 polarity of biochar, and O/C indicates the content of oxygen functional groups in biochar [18].
11 Comparing the results before and after BB adsorption, the ratio of O/C and (O+N)/C decreases,
12 which indicates that the number of oxygen-containing functional groups in the biochar after
13 adsorption decreases, the hydrophilicity of the biochar surface decreases, and the polarity of the
14 biochar decreases. After the CB adsorption, the hydrophilicity increases, and the oxygen-
15 containing functional groups are almost unchanged. The H/C content of CB and BB increases
16 after adsorption, which indicates that the aromaticity of the two biochars weakened after
17 adsorption. The ratio of O/C and (O + N)/C of BB decreases, which may be caused by a large
18 amount of removal of -OH and -COO in the process of dye adsorption, resulting in the decrease of
19 O content. However, the slight increase of O/C and (O + N)/C ratios of CB may be due to
20 irreversible adsorption of dye MG molecules, and the specific surface area and pore volume of
21 CB are obviously larger than those of BB. BB has significantly more oxygen-containing
22 functional groups than CB. It is speculated that the adsorption of CB for MG is mainly based on

1 physical effects and supplemented by chemical effects. The adsorption of BB for MG is mainly
2 due to the chemical adsorption by the oxygen-containing functional groups, followed by physical
3 adsorption. These results also explain the much higher adsorption capacity of BB than that of CB.

4 5 3.1.3. TG-DTG analysis

6 Fig. 2 shows the TGA results for the adsorbents. The thermal degradation behaviour of the two
7 biochars was studied between 35-1,005°C. On the whole, the pyrolysis laws of the two biochars
8 are roughly identical, but BB has faster degradation rate than CB. The TG/DTG curve can be
9 divided into three stages. The first stage (35-200°C) is caused by the evaporation of water
10 absorbed by BB and CB as free bound water, which corresponds to DTG weight loss peaks at
11 67.526°C, 157.32°C and 77.407°C, respectively. The second stage (200-450°C) is the thermal
12 degradation of cellulose, hemicellulose and other volatile molecules [19]. The third stage (450-
13 1,005°C) is mainly due to the combustion of residual lignin and the formation of coke. At this
14 temperature, CO and CO₂ may be formed due to the oxidation of non-volatile carbon molecules
15 [20]. The total weight loss of BB and CB is 42.5% and 42.1%, respectively.

16 17 3.1.4. BET analyses

18 Fig. 3 shows four adsorption analysis curves and pore size distribution curves, which correspond
19 to BB and CB before and after adsorption. As shown in Table 3, after adsorption, the specific
20 surface area and pore volume of the two biochars decrease, and the pore size doubles. Two
21 reasons are speculated: on one hand, MG fills the pores of the adsorbents due to physical
22 adsorption; on the other hand, the functional groups on the pore surface react with MG to

1 generate new substances. In addition, Fig. 3(b) shows that the pores of CB and BB are mainly
2 mesopores of 2-40 nm [21]. Fig. 3(a) shows the analytical N₂ adsorption curves of the two
3 materials before and after adsorption. The N₂ adsorption curves belong to type IV. The nitrogen
4 adsorption and desorption curves of all materials are different, and all curves of the four
5 materials show hysteresis. The BB hysteresis band range is $0.5 < p/p_0 < 1$, the CB hysteresis
6 band range is $0.1 < p/p_0 < 0.9$, and the hysteresis curves of the two materials are retained after
7 adsorption. The appearance of hysteresis loops is due to the porosity of the materials, which
8 leads to capillary condensation systems. The end of hysteresis indicates that the mesopores of the
9 material are filled, and the adsorption isotherm continues to rise because the adsorbate molecules
10 interact and continue to adsorb to form a multimolecular layer [22].

11

12 3.1.5. FTIR analysis

13 Fig. 4 shows the infrared spectra of BB and CB before and after the MG adsorption. The peaks
14 of BB and CB at $2,920 \text{ cm}^{-1}$ and $2,922 \text{ cm}^{-1}$ belong to the strong aliphatic C-H stretching band
15 [23]. The BB and CB peaks at $2,920 \text{ cm}^{-1}$ and $2,922 \text{ cm}^{-1}$ belong to the broad band of hydroxyl (-
16 OH) groups at $3,000\text{-}3,500 \text{ cm}^{-1}$ and $3,500 \text{ cm}^{-1}$, respectively. After adsorption, the peak value of
17 BB increases, and the peak value of CB disappears. The peak at approximately $1,600 \text{ cm}^{-1}$ can be
18 attributed to the stretching vibration of C=C bonds in aromatic groups [24] and is characterized
19 by the enhancement and disappearance after adsorption for CB and BB, respectively. For BB, the
20 peak at $1,000\text{-}1,200 \text{ cm}^{-1}$ is attributed to C-O tensile vibrations [25], and the peak disappears
21 after adsorption. The peak range of $800\text{-}600 \text{ cm}^{-1}$ represents the C-H rocking vibration of
22 heteroaromatic and aromatic compounds and various alkyl halides [26]. Peaks below 600 cm^{-1}

1 belong to M-X (M: metal; X: halogen) in organic and inorganic halogen compounds, and the
2 peak shift from 517 cm^{-1} before adsorption to 576 cm^{-1} after adsorption for BB indicates that BB
3 interacts with the dye organic functional groups after the MG adsorption [27]. The results show
4 that the aromatic structure of the two types of biochar is mainly formed due to the condensation
5 structure of cellulose, hemicellulose, lignin and protein in the biochar during pyrolysis [28].
6 Moreover, a large number of organic active groups (hydroxyl, carboxyl, carbonyl (-OH, CO, CH)
7 and benzene rings) can provide electrons for the dye, and the electrostatic attraction between
8 aromatic hydrocarbons in BB and CB and the aromatic structure of MG promotes the dye
9 adsorption by the adsorbents [29].

10

11 3.1.6. XRD analysis

12 Fig. 5 shows the XRD analysis results of BB and CB. The XRD patterns of BB and CB before
13 and after adsorption are shown in the figure. The crystal substance corresponding to BB is
14 KHCO_3 , which is consistent with the powder diffraction file (PDF) database (PDF card: 73-
15 2155). After adsorption, the crystal structure is destroyed, and a new substance, $\text{C}_{14}\text{H}_{12}\text{O}_4$, is
16 formed, which is consistent with the PDF database (PDF card: 50-2270). MG may react with the
17 hydroxyl, carboxyl and carbonyl groups of organic active functional groups in BB to produce
18 new substances [30]. CB has two wide diffraction peaks with no formed crystal structure, which
19 correspond to angles of 42° and 30° . After adsorption, the diffraction peaks become smooth, and
20 the overall structure has no obvious change.

21

22 3.1.7. TOC analysis

1 The TOC of the solution before and after adsorption was determined by TOC analyzer. The
2 results showed that the TOC content of the original solution was 287.7 mg/L, the TOC content of
3 the solution after BB adsorption was 30.63 mg/L, and the TOC content of the solution after CB
4 adsorption was 63.70 mg/L, which indicated that MG and its derivatives had indeed been
5 removed.

6

7 **3.2. Adsorption Experimental Analysis**

8 3.2.1. Effect of dosage

9 As shown in Fig. S2, when the dosage of the two adsorbents gradually increases from 0.5 g/L to
10 3.5 g/L, the adsorption amount first increases and subsequently decreases, and the optimal
11 dosage of the two biochars is 2 g/L. The adsorption amount is limited, and a moderate increase in
12 amount of adsorbent is beneficial to the adsorption of MG by BB and CB. This effect occurs
13 because with the increase in adsorbent dose, the adsorption surface area and number of
14 functional groups in the adsorption increase [31]. Meanwhile, the equilibrium concentration
15 decreases with the increase in addition amount. According to the adsorption equilibrium law, the
16 adsorption capacity decreases.

17

18 3.2.2. Effect of different pH values

19 pH is one of the main factors that determine the solution adsorption process [32]. Thus, the pH of
20 the solution greatly affects the surface properties of biochar. The left coordinate axis of Fig. S3
21 shows the effect of the initial pH on the adsorption amount. When the pH increases, the
22 adsorption of MG by the adsorbents first increases and subsequently stabilizes. At pH = 5, the

1 adsorption capacities of BB to MG and CB to MG were 936.02 mg/g and 249.91 mg/g,
2 respectively. The right coordinate axis of Fig. 3 shows the pH change in the solution after
3 adsorption. After adsorption, the pH of the solution increases relative to the initial pH. The zero
4 charge results show that $BBpzc = 9.84$, $CBpzc = 10.26$, and the $pHpzc$ of both biochar is greater
5 than 8. When $pH < pHpzc$, the surface of biochar is positively charged and forms electrostatic
6 repulsion with cationic dye MG, which leads to the inhibition of the adsorption capacity of
7 biochar [33]. Therefore, this may be the reason for the poor adsorption effect of BB and CB at
8 low pH. With the increase of pH, the negative charge on the surface of biochar increases, the free
9 H^+ in the solution gradually decreases, and the electrostatic repulsion weakens, making the MG
10 adsorption effect gradually enhance [34]. However, when the pH is further increased from 4 to 8,
11 the adsorption capacity basically tends to be stable without obvious change, indicating that the
12 adsorption mechanism of BB and CB to MG may have other adsorption mechanisms besides
13 electrostatic adsorption.

14

15 3.2.3. Adsorption kinetics

16 The pH of the solutions containing BB and CB was adjusted to 4 and 5, respectively, to study the
17 effect of different reaction times on the adsorption of MG by the two adsorbents. The results are
18 shown in Fig. S4. At different concentrations, BB adsorbs MG with a fast phase (0-30 min), a
19 medium phase (30-60 min) and a slow phase (60-480 min). CB at a low concentration of 150
20 mg/L quickly reaches adsorption equilibrium, and at a concentration of 250 mg/L, CB adsorption
21 of MG can be divided into fast (0-120 min) and slow phases (120-480 min). At a concentration
22 of 500 mg/L, CB adsorption of MG can be divided into fast (0-60 min), medium (60-240 min)

1 and slow phases (240-480 min). Thus, both adsorbents can basically reach adsorption
2 equilibrium within 1 h, which exhibits high adsorption efficiency and great practical application
3 potential. However, to ensure that adsorption equilibrium can be reached under all conditions, 4
4 h was taken as the adsorption equilibrium reaction time.

5 Pseudo-first-order kinetics (Eq. (2)) and pseudo-second-order kinetics (Eq. (3)) models
6 were used to fit the experimental data, and intraparticle diffusion was analysed according to Eq.
7 (4).

$$8 \quad Q_t = Q_e(1 - e^{-K_1 t}) \quad (2)$$

$$9 \quad Q_t = Q_e t(t + 1/k_2 Q_e) \quad (3)$$

$$10 \quad Q_t = K_3 t^{1/2} + C \quad (4)$$

11 where Q_t (mg/g) is the adsorption amount at time t ; Q_e (mg/g) is the equilibrium adsorption
12 amount; t (min) is time; K_1 (1/min) is the rate constant.

13 Table S1 presents the fitting results and shows that the pseudo-second-order kinetic
14 correlation coefficient R^2 is greater than the pseudo-first-order kinetic correlation coefficient
15 under different concentration conditions. This result shows that the pseudo-second-order kinetic
16 model is more consistent with the entire process of MG adsorption by biochar, so the adsorption
17 of MG by BB and CB involves both pure physical adsorption and chemical adsorption such as
18 electron transfer and exchange. In addition, BB has significantly smaller K_1 value than CB,
19 which indicates that CB has a higher affinity than BB for negatively charged dyes. The
20 correlation coefficients of the intraparticle diffusion model indicate that there is no linear
21 relationship for the adsorption of MG by BB and CB (the correlation coefficients are 0.6274,

1 0.7856, 0.8610, 0.5733, 0.7521, and 0.7779), which shows that the adsorption of MG is not only
2 controlled by intraparticle diffusion. Thus, chemisorption is the rate-controlling step of BB and
3 CB adsorption of MG; chemisorption may include the external liquid film diffusion, surface
4 adsorption and intraparticle diffusion [35].

5

6 3.2.4. Adsorption isotherms

7 As shown in Fig. S5, the adsorption equilibrium curves of the two adsorbents for MG are slightly
8 different. In this paper, two typical isothermal adsorption models (Langmuir and Freundlich)
9 were used to fit the data to study the distribution of adsorbed molecules in the solid and liquid
10 phases at equilibrium. Eq. (5) and Eq. (6) provide the fitting equations for the two models:

$$11 \quad Q_e = K_F C_e^{1/n} \quad (5)$$

$$12 \quad Q_e = \frac{C_e K_L Q_m}{1 + K_L C_e} \quad (6)$$

13 where Q_e (mg/g) is the equilibrium adsorption amount; C_e (mg/L) is the MG concentration after
14 adsorption; K_L (L/mg) is Langmuir equilibrium constant; Q_m (mg/g) is the theoretical saturated
15 adsorption amount; K_F and n are Freundlich constants.

16 The fitting of the two models at room temperature is shown in Fig. S5, and the resulting
17 fitting parameters are shown in Table S2. The results show that both isotherm models well fit the
18 adsorption process of MG by the two adsorbents. However, the correlation coefficient of the
19 Freundlich model is higher ($R^2 > 0.9100$), which shows that the adsorption process of the two
20 biochars is more consistent with Freundlich model, and the interactions between adsorbed
21 molecules involve multilayer adsorption. For both adsorbents, $1/n < 1$, which indicates that MG

1 can be well adsorbed by BB and CB; the adsorption effect of BB is better. Comparing K_L and K_F
2 values, we find that BB has better adsorption and binding stability for MG than CB [36].

3 Table 2 shows the physical properties and MG adsorption capacity of different adsorbents.
4 In terms of preparation technology, this study is characterized by simple preparation and a low
5 risk of secondary pollution. Although BB and CB have slightly poor physical properties, their
6 adsorption of MG has obvious advantages. As a whole, biochar is an inexpensive and readily
7 available adsorbent material that can provide reference value for the MG wastewater treatment.

8

9 **3.3. Adsorption Mechanism**

10 In this study, the adsorption of MG by the adsorbents BB and CB was studied. The adsorption
11 experiment and characterization results imply that the adsorption mechanism of CB and BB
12 mainly includes chemical adsorption and physical adsorption. The adsorption mechanism is
13 shown in Fig. S1. The physical effect is mainly pore filling. The SEM results and N_2 adsorption-
14 desorption isotherms show that CB and BB have a hollow structure. By reducing the steric
15 hindrance effect, the physical adsorption capacity of an adsorbent can be increased. Regarding
16 chemical interactions, FTIR and XRD analysis confirm that the biochars have electronic
17 functional groups such as hydroxyl, carboxyl, and carbonyl groups, and these organic active
18 groups can provide π electrons to promote the MG adsorption, and π - π interactions and cation- π
19 interactions by the biochar.

20

21 **4. Conclusions**

1 This study proves that banana straw and cassava straw are promising and can be used to produce
2 low-cost biochar to remove MG. The maximum adsorption capacities of the biochars can reach
3 1,092.80 mg/g and 261.42 mg/g, respectively. The characterization analysis shows that BB and
4 CB have large specific surface area and pore size, and the biochars are rich in oxygen-containing
5 functional groups. These characteristics provide the biochars with a strong adsorption capacity.
6 Furthermore, the temperature and pH play an important role in the MG adsorption, which can be
7 due to the electrostatic repulsion between positive charges. Adsorption by the biochars conforms
8 to Freundlich model, and kinetic studies show that the reaction with MG conforms to the pseudo-
9 second-order kinetic model. This technology to convert waste biomass to biochar will be of great
10 significance for maximizing the adsorption of highly toxic textile dyes.

11

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16

17 **Author Contributions**

18 D.H. (Professor) and H.L. (Professor) provided research ideas and designed the experiments. Y.L.
19 (postgraduate student) conducted the experiments and wrote and revised the manuscript. T.S.
20 (student) mainly conducted experiments and collected and sorted the data. L.A. (Ph.D. student)
21 read the manuscript and suggested modifications.

22

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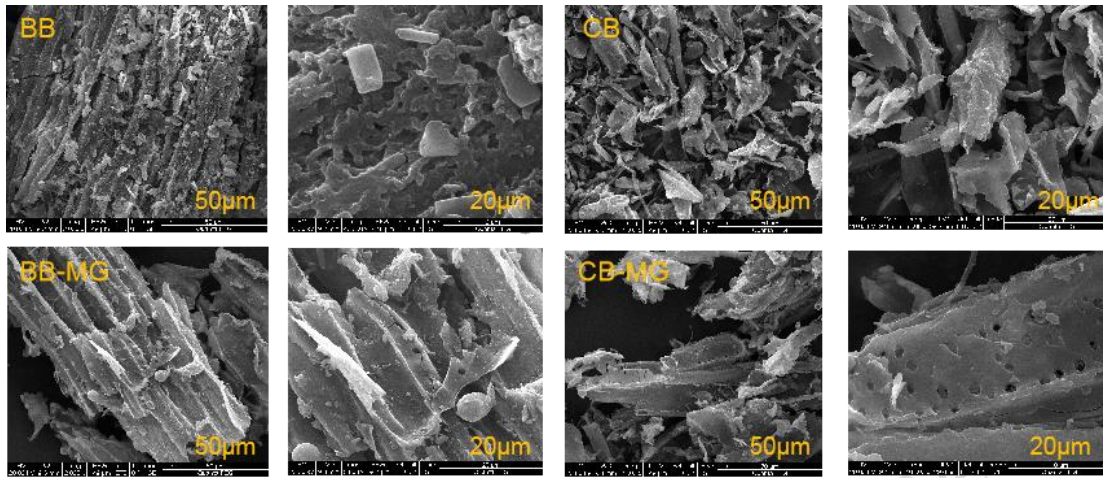


Fig. 1. SEM-EDS analysis of BB and CB before and after MG adsorption.

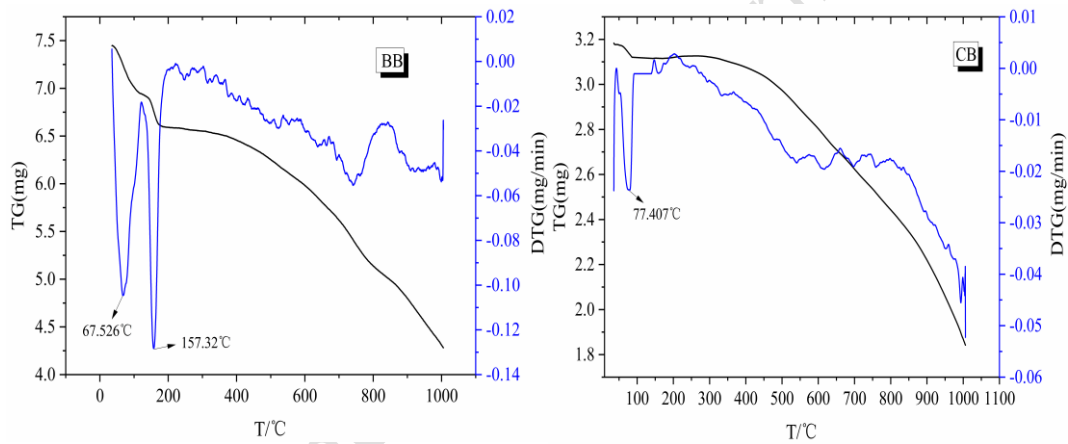


Fig. 2. TG / DTG curves of BB and CB.

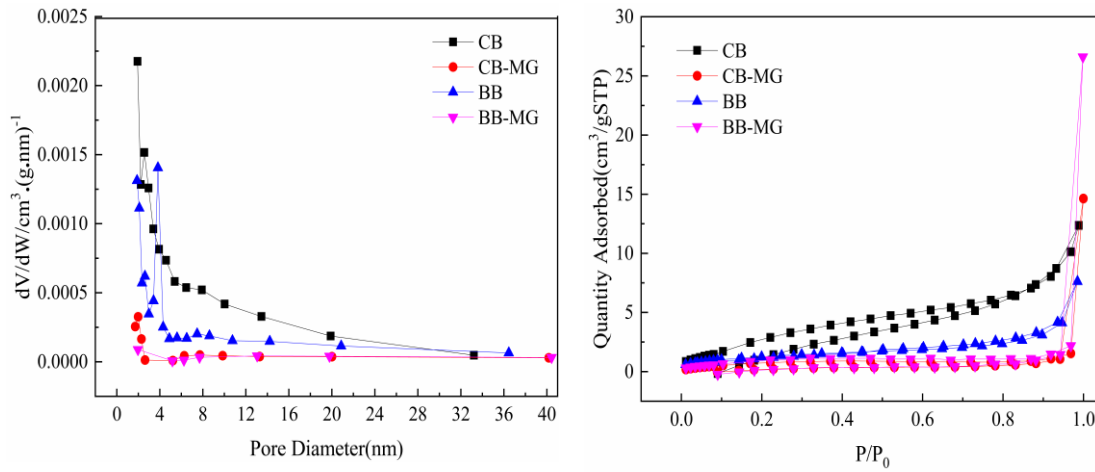


Fig. 3. Biochar daughter-in-law analysis curve.

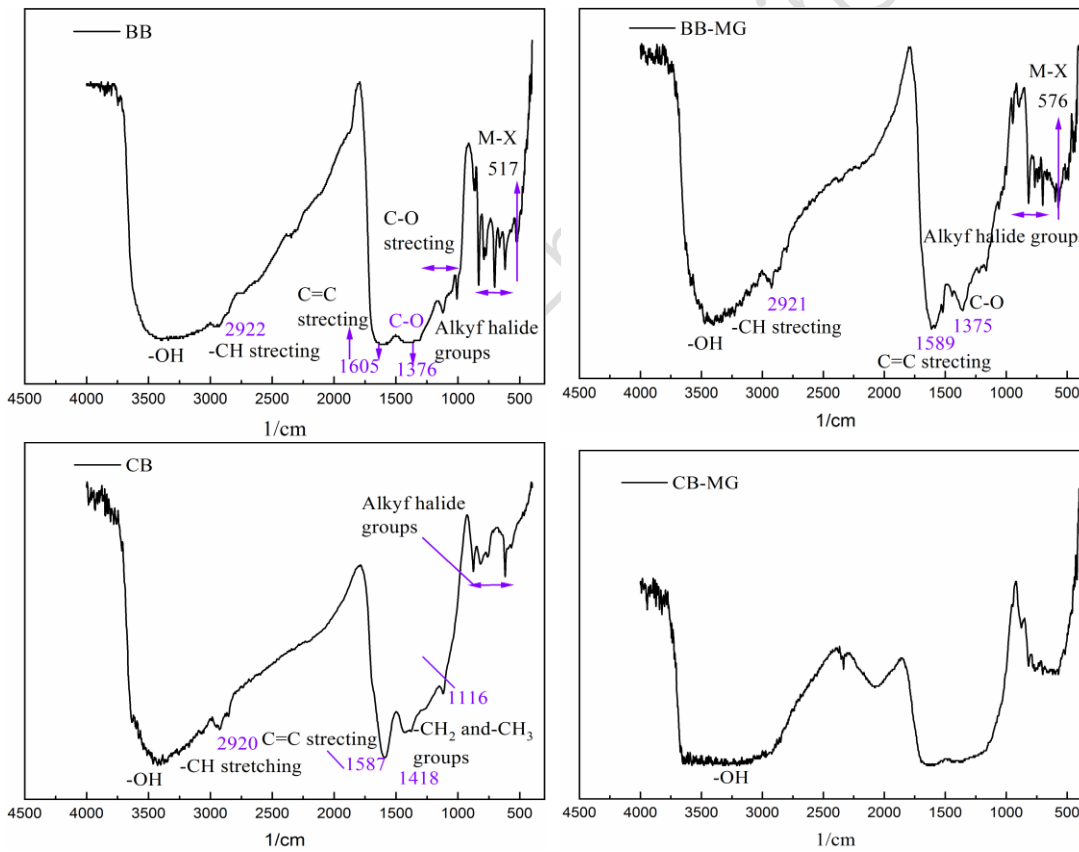


Fig. 4. FTIR spectra before and after BB and CB adsorption.

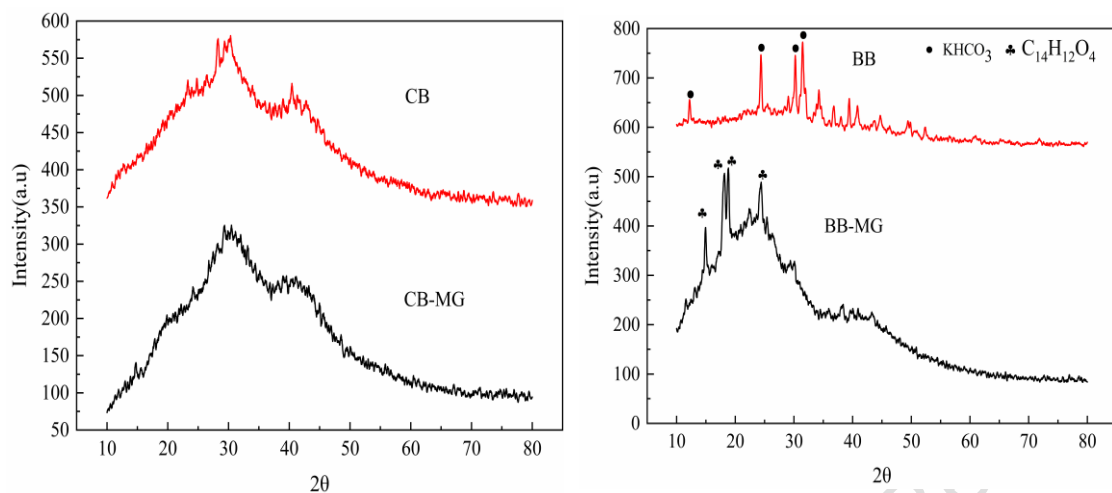


Fig. 5. XRD analysis chart of BB and CB.

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Table 1. Element Analysis Table

	C(%)	H(%)	N(%)	O(%)	Ash Content(%)	H/C(%)	O/C(%)	(O+N)/C(%)
CB	69.52	3.412	2.26	16.248	8.56	0.049	0.234	0.266
CB-MG	73.56	3.656	2.555	18.203	2.026	0.050	0.247	0.282
BB	43.96	2.705	2.898	47.092	3.345	0.062	1.071	1.137
BB-MG	66.685	4.37	4.336	22.803	1.806	0.066	0.342	0.407

1 **Table 2.** Studies on Biochar Adsorption of MG

Adsorbent	Tk	dosage g/L	Co mg/L	Qe mg/g	pH	Th	Adsorption Isotherm	BET (m ² /g)	Total pore volume (cm ³ /g)	Average pore diamete (nm)	Ref.
Preparation of nanochitosan-STP from shrimp shel	298	0.3	10-500	317.73	6	3	Freundlich /Langmuir	-	-	-	[12]
activated charcoal of Castor bean pericarp	300	20	25-200	27.78	7	1	Langmuir	-	-	-	[1]
activated charcoal of Palm flowe	303	10	25-100	48.48	6.87	4	Langmuir	9.52	0.0737	246.26	[21]
activated carbon of bamboo	303	1	25-300	263.58	4	1	Langmuir	1724	1.073	2.485	[4]
Bagasse activated carbon	303	0.3	15-100	13.2	8	1.923	Freundlich	82.23	0.0253	-	[20]
magnetic Cortaderia selloana flower spikes	298	4	25-350	56.5	6	0.75	Langmuir	-	-	-	[7]
Oyster mushroom bio-composite material	298	5	50-150	77.11	7	1.5	Langmuir	-	-	-	[26]
natural red clay	313	8	100-250	84.75	5.16	0.75	Freundlich	-	-	-	[28]
Magnetic Nickel Oxide Nanoparticles	-	-	50	87.72	7	0.66	Langmuir	11.617	0.0183	1.22	[29]
Opuntia ficus-indica Activated biochar	303	0.6	100- 1,000	1,341.38	6	40	Langmuir/S ips/Hill	0.093	11.51	0.088	[15]
Epoxy triazine triketone functionalized magnetic nanoparticles	303	5	-	84.78	10	0.5	Freundlich	847.88	-	0.025~0.052	[11]
Magnetic Cobalt Oxide Nanoparticles	298	1	60	238.1	7	2	Langmuir	8.351	0.0108	0.711	[12]
SBA-15 mesoporous nanoparticles	298	0.1	50-70	219	9	0.233	Langmuir	602.5	0.69	4.58	[13]

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1 **Table 3.** Biochar BET Pore Size and Pore Volume Parameters

Biochar	CB	CB-MG	BB	BB-MG
BET (m ² /g)	13.2564	3.0619	4.6708	3.7749
Average pore diameter (nm)	5.33087	12.7169	9.4845	18.8421
Total pore volume (cm ³ /g)	0.01766	0.0097	0.01107	0.0178

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